Respiration rates in marine heterotrophic bacteria relate to the cytometric characteristics of bacterioplankton communities

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Respiration in marine bacteria (fraction $< 0.6 \ \mu$ m) was investigated in 5 distinct coastal ecosystems. Respiration rates were estimated from the decrease of dissolved oxygen (O₂) concentrations measured with a fibre-optic oxygen sensor (optode) during short-term incubations (5 to 6 hours) in the dark. Sub-populations of heterotrophic bacteria were defined and enumerated by flow cytometric analysis. Respiration rates ranged from 0.04 to 0.14 μ mol O₂ l⁻¹ h⁻¹ and were not correlated to total bacteria abundances. In contrast, they significantly increased with the relative proportion of HDNA cells (i.e. high deoxyribonucleic acid content) within the bacterioplankton community. These results stress the need to relate bacterial respiration rates to the absolute and relative abundance of the different cytometrically defined sub-populations found in bacterial assemblages, rather than to total cell counts. This result indicates that only a fraction of bacterial cells may contribute to bacterial respiration and the related biogeochemical fluxes may be required.

Keywords: marine bacteria, respiration rates, flow cytometry, bacteria abundances, community structure

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INTRODUCTION

Heterotrophic bacteria are a fundamental element in the ocean carbon cycling process (Cole et al., 1988). They contribute to nutrient and carbon cycling through the production of new bacterial biomass (secondary production) generally transferred to higher trophic levels through the microbial food web and the remineralization of organic carbon and nutrients (bacterial respiration) (del Giorgio & Cole, 1998). It is now established that 30-60% of phytoplankton production (i.e. the amount of inorganic carbon fixed by phytoplankton photosynthesis, corrected for phytoplankton respiration) is processed by bacteria in both marine and freshwater systems (Cole et al., 1988; Ducklow & Carlson, 1992). While planktonic heterotrophic bacteria are responsible for the bulk of microbial respiration (Williams, 1981; Cole et al., 1988; del Giorgio & Duarte, 2002), previous studies stressed the lack of any correlation between respiratory activity and quantitative descriptors of the marine bacterial community (Pomeroy et al., 1994; Briand et al., 2004). An improved knowledge and understanding of factors controlling respiration in marine bacteria is then critically needed to further assess the fate of carbon within aquatic ecosystems.

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Bacterial communities are intrinsically made of complex assemblages of numerous species having different metabolic rates (Servais et al., 2001). The relationship between respiration rates and abundances has nevertheless previously been investigated considering the heterotrophic bacteria community as a whole. Flow cytometry allows for the differentiation of subpopulations of bacteria via the light-scattering properties (related to size) and the DNA (deoxyribonucleic acid) content of individual cells (Li et al., 1995; Gasol & del Giorgio, 2000; Lebaron et al., 2002; Servais et al., 2003). Generally, 2 main groups, represented by cells with high (HDNA) and low deoxyribonucleic acid content (LDNA), are observed in aquatic samples (Lebaron et al., 2002; Servais et al., 2003; Bouvier et al., 2007). Variations in DNA content have been related to single cell activity with HDNA and LDNA cells constituting respectively the active cells and the dormant or dead cells (Li et al., 1995; Gasol & del Giorgio, 2000; Lebaron et al., 2002; Servais et al., 2003). The percentage of HDNA bacteria has widely been considered as an accurate indication of the proportion of active cells within microbial communities (Jellett et al., 1996; Gasol et al., 1999). However, recent studies do not support the simple dichotomy of HDNA as active and LDNA as inactive cells (Bouvier et al., 2007) as there is now evidence that low DNA cells may also be active (e.g. Zubkov et al., 2001; Longnecker et al., 2005). Although these two distinct fractions of bacterioplankton have been observed in a wide range of environments (e.g. Bouvier *et al.*, 2007), little is still known concerning their metabolic and ecological role in marine ecosystems.

Respiration rates in heterotrophic bacteria are most often estimated by measuring changes of dissolved oxygen concentration during in situ or laboratory incubations, typically over 24 hours or more (e.g. Pomeroy et al., 1994, 1995). Since it has been shown that changes of bacterial community composition (Massana et al., 2001; Gattuso et al., 2002) and/or exhaustion of trophic resources (Pradeep Ram et al., 2003) are likely to occur during these long-term incubations, reducing the incubation times has become an absolute prerequisite to prevent the acquisition of results that may not be representative of the natural bacterial community. However, short-term incubations require the use of highly sensitive methods to obtain reliable data, as the rate of respiration of the microbial community is often very low. Oxygen microprobes (i.e. Clark electrode) or a type of optical electrodes, the so-called opt(r)odes, are more sensitive than the conventional method of Winkler and constitute reliable techniques to estimate the consumption of dissolved oxygen during an incubation time less than 24 hours (Briand et al., 2004; Warkentin et al., 2007). In addition to allowing for reliable respiration rate measurements even at low consumption rates, optodes present many advantages compared to microprobes (Warkentin et al., 2007): (i) they are used in non-invasive systems; (ii) no oxygen is consumed by the optode itself; and (iii) no mechanical stress, i.e. constant stirring of the samples, has to be imposed on bacteria. This technology has been recently used to estimate bacterial respiration in freshwater (Warkentin et al., 2007). However, to our knowledge, this technique has until now, never been applied for the estimation of bacterial respiration in marine ecosystems.

In this context, we combined a short-term incubation procedure based on the use of optodes coupled to flow cytometric analyses (FCM) to investigate the potential relationships between respiration rates and the abundances of flow cytometrically-defined sub-populations of heterotrophic bacteria. These experiments were carried out in 5 distinct ecosystems representative of typical coastal environments, i.e. seagrass bed, mudflat, sandflat, rocky shore and sheltered marina. These environments are likely to present qualitatively and quantitatively different communities of heterotrophic bacteria. The objective of this research was to investigate how the respiration rates in marine bacteria were related to the cytometric structure of the bacterioplankton communities and to provide new insight into the metabolic and ecological role of the cytometrically defined sub-populations of bacteria in marine ecosystems.

MATERIALS AND METHODS

Study sites and sampling

Bacterial respiration was investigated at several coastal sites, scattered along the coasts of two contrasted inlets on the southern coast of Australia, the Spencer Gulf and the Gulf Saint Vincent (Figure 1), both microtidal systems (i.e. tidal range < 2 m). Spencer Gulf extends 300 km northward from the continental margin of southern Australia. The coastal geometry affords some shelter from predominantly south-west ocean-generated swell (Fuller *et al.*, 1994; Harris, 1994; Porter-Smith *et al.*, 2004). This coastal region receives negligible terrigenous inputs as a consequence of a semiarid climate (annual rainfall < 550 mm), low relief and small riverine discharge (Fuller *et al.*, 1994; Schwarz, 2003). In contrast, the shallow and sheltered Gulf Saint Vincent is impacted by stormwater discharges, particularly near metropolitan Adelaide (Edyvane, 1996, 1999). The study sites (Figure 1; Table 1) have



Fig. 1. Study area and location of the 5 sampling sites (\$\priv:). (1) Sheltered marina (Adelaide area); (2) seagrass bed (Port Lincoln Jetty); (3) sandflat; (4) mudflat; (5) rocky shore (Port Lincoln area).

		from the 5	sampling sites (average \pm SD calculated o	n 3 samples). Fo	r the marina HDN	A abundances represent th	the sum of HDNA1 and HD	NA2.	A dame to e
Stations	S	(O°) T	NH_4^+ ($\mu\mathrm{M}$)	$NO_3^- + NO_2^- (\mu M)$	PO_4^{3-} (μM)	Chl a ($\mu g l^{-1}$)	Total (ml ⁻¹)	LDNA (ml ⁻¹)	HDNA (ml ⁻¹)	WDNA %
Rocky shore (5^*)	41.0	23.1	7.1	<dl< td=""><td><dl< td=""><td>0.16</td><td>$1.6\times10^{6}\pm1.5\times10^{5}$</td><td>$6.2\times10^5\pm5.1\times10^4$</td><td>$1.1\times10^{6}\pm6.4\times10^{3}$</td><td>$61.9\pm1.2$</td></dl<></td></dl<>	<dl< td=""><td>0.16</td><td>$1.6\times10^{6}\pm1.5\times10^{5}$</td><td>$6.2\times10^5\pm5.1\times10^4$</td><td>$1.1\times10^{6}\pm6.4\times10^{3}$</td><td>$61.9\pm1.2$</td></dl<>	0.16	$1.6\times10^{6}\pm1.5\times10^{5}$	$6.2\times10^5\pm5.1\times10^4$	$1.1\times10^{6}\pm6.4\times10^{3}$	61.9 ± 1.2
Mudflat (4 [*])	37.5	25.6	11.9	4.7	3.2	0.11	$1.6 imes 10^6 \pm 1.1 imes 10^5$	$2.6 \times 10^5 \pm 6.9 \times 10^4$	$1.3 imes 10^6 \pm 1.2 imes 10^5$	83.8 ± 4.4
Seagrass bed (2 [*])	37.1	20.5	7.1	1.6	<dl< td=""><td>0.19</td><td>$1.5 \times 10^{6} \pm 4.4 \times 10^{4}$</td><td>$4.4 \times 10^5 \pm 9.7 \times 10^4$</td><td>$1.0 \times 10^{6} \pm 5.3 \times 10^{4}$</td><td>$70.3 \pm 5.7$</td></dl<>	0.19	$1.5 \times 10^{6} \pm 4.4 \times 10^{4}$	$4.4 \times 10^5 \pm 9.7 \times 10^4$	$1.0 \times 10^{6} \pm 5.3 \times 10^{4}$	70.3 ± 5.7
Sheltered marina (1*)	37.0	22.0	8.9	1.6	1.1	0.17	1.3×10^{6}	1.8×10^5	1.1×10^{6}	85.5
Sandflat (3 [*])	36.0	24.1	5.4	1.6	2.1	0.27	$1.1 imes 10^6 \pm 2.3 imes 10^5$	$1.5 \times 10^5 \pm 3.9 \times 10^4$	$9.1 \times 10^5 \pm 2.6 \times 10^5$	85.4 ± 5.6

corresponding numbers on Figure 1.

Table 1. Salinity (S), temperature (T; °C), ammonium (NH_{+}^{+}), nitrite + nitrate ($NO_{3}^{-} + NO_{2}^{-}$), phosphate ($PO_{3}^{2}^{-}$) concentrations (μ M) and chlorophyll-a (Chl a) concentrations (μ g I⁻¹) measured at the 5 sampling

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thus been chosen because they are representative of the typical coastal environments found along the Spencer Gulf (i.e. seagrass beds, mudflats, sandflats and rocky shores) and the Gulf Saint Vincent (i.e. sheltered marina).

Samples were collected between November and December 2006 (Table 1). At each site, temperature and salinity were recorded using a Hydrolab Data Sonde 3 (Hach[®], USA). Water samples for environmental parameters (nutrients and chlorophyll-*a*), heterotrophic bacteria abundance and bacterial respiration were carefully collected from sub-surface (10 cm depth) in 50 cm of water, close to the shore line, using an acid-washed 1-l borosilicate bottle.

Environmental parameters

Dissolved inorganic nutrients concentrations were determined from 12 ml filtered (Whatman GF/C) water samples. Analyses were performed in the field using a portable LF 2400 photometer (Aquaspex[®], Australia) according to standard colorimetric methods for NH_4^+ (indophenol blue), NO_2^- (naphtylethylene diamine), NO_3^- (naphtylethylene diamine after zinc reduction) and PO_4^{3-} (ascorbic acid reduction). The limit of detection was 0.6 μ M for ammonium, 0.2 μ M for nitrite, 1.6 μ M for nitrate and 1.1 μ M for phosphate.

Samples (50 ml) for chlorophyll-*a* (Chl *a*) extraction were filtered through glass-fibre filters (Whatman GF/C) and immediately frozen (-20° C) until analysis. Chlorophyllous pigments were then extracted in 5 ml of methanol in the dark at 4° C during 24 hours (Herbland *et al.*, 1985), and concentration of Chl *a* (µg l⁻¹) was determined following Strickland & Parson (1972) using a Turner 450 fluorometer (Sparks[®], USA) previously calibrated with a pure Chl *a* solution (*Anacystis nidulans* extract, Sigma Chemicals, St Louis).

Heterotrophic bacteria abundance

Heterotrophic bacteria were identified and enumerated using a FACScantoTM flow cytometer (Becton-Dickinson, USA). Samples $(3 \times 1 \text{ ml})$, fixed with 0.5% (final concentration) of glutaraldehyde in the dark at 0°C for 15 minutes, quick frozen in liquid nitrogen and then stored at -80°C. After being quick thawed, samples were diluted (1:10) in 0.2 µm filtered TE buffer (10 mM Tris, 1 mM EDTA, [pH 8]), stained with SYBR-I Green solution (1:5000 dilution) and incubated at 80° C in the dark for 10 minutes. Fluorescent beads of 1 μ m diameter (Molecular Probes, Eugene, Oregon) were added to all samples, as size and concentration standards. Working beads concentrations were estimated after each session under epifluorescence microscopy, to ensure reliability of the beads concentration (Gasol & del Giorgio, 2000) and all FCM parameters were normalized to bead concentration and fluorescence. For each sample, forward-angle light scatter (FSC), side-angle light scatter (SSC), green (SYBR-I) fluorescence, red (chlorophyll) fluorescence, and orange (phycoerythrin) fluorescence were acquired. Populations of heterotrophic bacteria were identified and quantified using the flow cytometry analysis software WinMDI 2.9 ([©]Joseph Trotter, USA). Discrete sub-populations of bacteria with high and low deoxyribonucleic acid content (HDNA and LDNA, respectively) were discriminated based on their differences in SYBR-I Green fluorescence and SSC, i.e. as non-overlapping classes of size and green fluorescence according to Lebaron et al. (2002) and Servais et al. (2003). These populations were identified as non-phototrophic bacterial

populations as their size and nucleic acid content were consistent with bacterial-sized organisms, and because they lacked red or orange fluorescence (indicative of chlorophyll or photopigment content; Marie *et al.*, 1999).

Bacterial respiration

Bacterial respiration was estimated from the evolution of oxygen concentrations in biological oxygen demand (BOD) glass bottles. Oxygen concentration was measured using a fibre-optic oxygen sensor, FIBOX 3 (PreSens® GmbH, Germany), equipped with a temperature compensation system. Planar oxygen sensors (i.e. optodes), based on a polyester support and optically isolated, were glued at the bottom of each BOD bottle. This method is based on the quenching of luminescence caused by collision between molecular oxygen and luminescent dye molecules in the excited state; after excitation with a blue light (excitation wavelength: 505 nm) the oxygen sensor emitted a red luminescent light (Kautsky, 1939), which was read out non-invasively through the transparent bottom of the bottle using an oxygen sensor probe. The life time of this red luminescence light depends on the ambient oxygen concentration.

Water samples were pre-screened through 250 μm mesh to remove mesozooplankton and subsequently gravity filtered through 0.6 µm pore size filters (Nucleopore), in order to remove most phytoplankton and protozoa (Williams, 1981; Briand et al., 2004; Warkentin et al., 2007). Samples were then carefully siphoned through a surgical-grade silicone tube in three overflowing 250 ml BOD dark bottles equipped with optodes. Sieving and filtration were performed with special care to avoid air bubbles. BOD bottles were then incubated in situ in sub-surface (10 cm depth). Incubations were carried out in the dark to avoid photosynthesis processes by potential remaining photoautotroph organisms. Oxygen concentration was measured every second in each BOD bottle during 1 minute and the concentration at a given incubation time was estimated by the mean value calculated from the resulting 60 measurements. Measurements were consistently performed 30 minutes after sampling and filtration, and every 30-60 minutes afterward over a 5 to 6 hours period.

Data analysis

The significance of the regression analyses was determined using Pearson's test calculated on 13 data points (3 bottles per site; 2 bottles were not considered because air bubbles were observed inside the bottles during the incubation). Multiple comparisons between sampling sites were conducted using the Kruskal–Wallis test (KW test), and a subsequent multiple comparison procedure based on the Tukey test was used to identify distinct groups of measurements (Zar, 1996).

RESULTS

Environmental parameters

Temperature and salinity varied from 20.5 to 25.6°C and from 36 to 41 (Table 1). Sampling sites were characterized by relatively high (NH_4^+) concentrations with values ranging from 5.4 to 11.9 μ M (Table 1). With 11.9 μ M for ammonium,

4.7 μ M for nitrate + nitrite and 3.2 μ M for phosphate, the mudflat site exhibited substantially higher nutrient concentration than those observed in the 4 others sites ($\leq 9 \mu$ M for NH₄⁺; $\leq 1.6 \mu$ M for (NO₃⁻ + NO₂⁻); ≤ 2.1 for PO₄³⁻). Chl *a* concentrations were low in the 5 sampling sites, with values typically smaller than 0.3 μ g l⁻¹ (Table 1).

Abundances of heterotrophic bacteria

No significant differences were observed in the abundances of total heterotrophic bacteria between sampling sites (KW test, P > 0.05) with values ranging from 1.1×10^6 to 1.6×10^6 ml⁻¹ at the sandflat and mudflat/rocky shore sites, respectively (Table 1). The bacterial communities were consistently characterized by a clear separation between HDNA and LDNA sub-populations (Figure 2). More specifically, the bacterial communities were characterized by one LDNA subpopulation and one HDNA sub-population, except in the sheltered marina where 2 distinct HDNA sub-populations were identified (Figure 2). LDNA and HDNA abundances ranged from 1.5×10^5 to $6.2\times10^5\,\text{ml}^{-1}$ and from $9.1\times$ 10^5 to 1.3×10^6 ml⁻¹, respectively (Table 1). At each sampling site the percentage of HDNA sub-populations was always greater than 50% and varied significantly between sampling locations (KW test, P < 0.05; Table 1). The fraction of HDNA cells was significantly higher at the sheltered marina, sandflat and mudflat sites with respectively 85.5, 85.4 and 83.8% of HDNA cells, than at the rocky shore site and over the seagrass bed where the percentage of HDNA cells remained below 71% (P < 0.05).

Respiration rates in heterotrophic bacteria

Two different trends of oxygen concentrations over time were identified (Figure 3), as previously observed and described in Briand *et al.* (2004). In two sites (marina and rocky shore), the slope was maximal at the beginning of the incubation and decreased thereafter (Figure 3A), whereas a decrease in oxygen concentration was only observed after a time lag of 150 to 250 minutes in the sandflat, mudflat and seagrass bed (Figure 3B). Consequently, an exponential decay was fitted to the part of the trend of oxygen concentration data versus time exhibiting the greatest slope (Figure 3) and respiration rates were then calculated from the first derivative of the fitted equation (Briand *et al.*, 2004).

Respiration rates exhibit significant variations between sampling sites (KW test, P < 0.05), with values ranging from 0.04 µmol O₂ l⁻¹ h⁻¹ over the rocky shore, to 0.14 µmol O₂ l⁻¹ h⁻¹ at the sheltered marina (Figure 4). The respiration rates obtained at the sheltered marina, at the mudflat and at the sandflat sites were significantly higher than those observed on the two other stations (P < 0.05). The respiration rates were not significantly correlated with total abundances of heterotrophic bacteria (N = 13; r = 0.33; P > 0.05) and HDNA abundances (N = 13; r = 0.37; P > 0.05). In contrast, the respiration rates were negatively correlated with LDNA abundances (N = 13; r = 0.70; P < 0.05) and positively correlated with % HDNA (N = 13; r = 0.72; P > 0.05; Figure 4). No significant correlations were observed between respiration rates and temperature and salinity (P > 0.05).



Fig. 2. Cytometric differentiation of heterotrophic bacterial populations in (A) the sandflat site (as a representative of all other sampling sites) and (B) in the sheltered marina. Histogram plots of green-fluorescence (SYBR Green) show peak related to LDNA (low DNA) observed in all sampling sites (A and B), and 1 (A) population or 2 (B) populations of increasing DNA content (HDNA1 and HDNA2). VLP, virus-like particles.

DISCUSSION

Long-term bottle incubations (i.e. more than 24 hours) have been widely used to estimate bacterial respiration in planktonic systems. However this technique presents a number of disadvantages that are now well documented, including changes in bacterial physiology (Malone et al., 1991), abundance and activity (del Giorgio & Cole, 1998), and community composition (Massana et al., 2001; Gattuso et al., 2002). It therefore appears fundamental to reduce the incubation duration to minimize the so-called 'bottle effect' and the concomitant community changes (Pomeroy et al., 1994; Gattuso et al., 2002; Briand et al., 2004). Nevertheless, the choice of incubation time relies on the adequate compromise between the limitation of the 'bottle effect' and the minimum time required to detect significant changes in oxygen concentrations. In the present study, incubations (5 to 6 hours) provided estimate of respiration rates of heterotrophic bacteria under conditions



Fig. 3. Two different temporal trends of oxygen concentrations $([O_2]; \mu \text{mol} \ l^{-1})$ observed during *in situ* incubations. (A) In the marina and the rocky shore sites, the slope was maximal at the beginning of the incubation and exponentially decreased thereafter. An exponential model was therefore fitted on the entire data set; (B) in the sandflat, mudflat and seagrass bed sites, oxygen decrease was only observed after a time lag, and the exponential decay was then fitted on the last part of the curve. The symbols are the mean values and the error bars are the standard errors calculated from the oxygen concentrations measured in the 3 BOD bottles for each incubation time.



Fig. 4. Bacterial respiration rates (BR, μ mol O₂ 1^{-1} h⁻¹) shown as a function of the percentages of HDNA cells (% HDNA) at the 5 sampling sites; mudflat, rocky shore, sheltered marina and seagrass bed and sandflat.

that remained as close as possible to the original *in situ* conditions. However, while pre-filtration procedures have been widely used to assess respiration rates in heterotrophic bacteria (see Table 2), this procedure could have consequences

References	Location	Filtration	Volume	Incubation	Data points	Method	BR (μ mol l ⁻¹ h ⁻¹)
Pomeroy et al., 1994	Northern Atlantic (21 locations)	No	125 ml	24 hours	5	Winkler	0.02-0.80
Gattuso et al., 2002	Alpine Lake (Switzerland)	1.2 µm	60 ml	55 hours	4	Winkler	0.00-0.15
Smith & Kemp, 2003	Chesapeake Bay (USA)	3.0 µm	300 ml	6 hours	2	Winkler	0.30-0.96
González et al., 2003	Bay of Biscay (Spain)	o.8 µm	130 ml	24 hours	2	Winkler	0.06-0.10
Briand <i>et al.</i> , 2004	Coastal lagoon (New Caledonia)	o.6 µm	250 ml	5 hours	Continuous	Microprobes	0.60-2.50
Kritzberg et al., 2005	10 Lakes (Wisconsin, USA)	2.7 µm	300 ml	24 hours	2	Winkler	0.06-1.37
Warkentin et al., 2007	River (Germany)	1.0 µm	3.5 ml	50 minutes	Continuous	Optodes	0.7-15.6
This study	5 coastal environments (South Australia)	o.6 µm	250 ml	5 to 6 hours	6 to 8	Optodes	0.04-0.14

Table 2. Literature review of (i) methods applied, i.e. filtration mesh size (μ m), incubation time (incubation), volume, data points and method used to estimate oxygen consumption and (ii) respiration rates (BR; μ mol l⁻¹ h⁻¹) estimated in different natural environments and in this study.

on the respiration rates observed. More particularly some bacteria attached to particles could have been retained on the filter, resulting in an under-estimation of the respiration rates observed in this study. However, given the lack of any procedure for blocking the metabolism of different sectors of the plankton with selective inhibitor size-fraction filtration procedure is a reliable technique to assess respiration rates in heterotrophic bacteria (e.g. Robinson & Williams, 2005).

Respiration rates in heterotrophic bacteria have mainly been estimated from the difference between initial and final oxygen concentrations, assuming a linear decrease of oxygen during incubation. Previous studies have nevertheless shown that decreases in oxygen concentration were not always linear during long-term (Biddanda et al., 1994; Pomeroy et al., 1994) and short-term incubations (Briand et al., 2004). In this study, two different trends of oxygen concentrations over time were identified (Figure 3). In two sites (marina and rocky shore), the slope was maximal at the beginning of the incubation and decreased thereafter (Figure 3A). This kinetic has been previously observed in rather eutrophic sites and could be explained by the exhaustion of trophic resources and subsequent decrease in bacterial activity (Briand et al., 2004). In contrast, the existence of a time lag before the observation of a significant decrease in oxygen, as observed in the sandflat, mudflat and seagrass bed (Figure 3B), was assigned to more oligotrophic sites where low net bacteria production was recorded (Briand et al., 2004). Even if in their study, Briand et al. (2004) were able to assign the different model to water of different trophic status, they have also underlined the difficulty to assign one model to any characteristic site. Future works are thus needed to link the different trend of oxygen consumption in heterotrophic bacteria to different trophic status.

The diversity of approaches (i.e. filtration, incubation time, number of data points and oxygen measurement techniques) used to estimate respiration rates in heterotrophic bacteria, through oxygen consumption, makes it difficult to compare the values observed in different ecosystems (Table 2). However, the respiration rates observed in the five coastal environments considered here are in the range of values reported previously for continental and oceanic waters (Table 2) and were significantly different between environments (Figure 4).

Bacterioplankton abundance and respiration rates

Sampling and incubations were conducted under different conditions of temperature, ranging from 20.5 to 25.6° C.

Temperature is known to profoundly and directly influence heterotrophic processes and metabolic activities in planktonic communities (Fenchel, 2005). However, we did not observe any correlation between temperature and respiration rates (P > 0.05), suggesting that temperature was not the main factor driving the spatial heterogeneity of respiration rates between our distinct sampling environments.

In a broad range of aquatic systems, bacterial production and respiration and bacterioplankton abundances have been shown to covary over a large range of temporal and spatial scales within a given system (Tuttle et al., 1987; del Giorgio et al., 1997; Bouvier et al., 2007). However, in our study respiration rates strongly fluctuated between sites (Figure 4) despite similar heterotrophic bacteria abundances and no significant correlation was observed between total bacterial abundances and respiration rates. Total bacteria abundances ranged from 1.1×10^{6} to $1.6 \times$ 10⁶ ml⁻¹ between sampling sites. These values are within the range of those previously described for eutrophic systems (i.e. 1 to 10×10^6 ml⁻¹; Cotner & Biddanda, 2002). This lack of significant correlation between total bacteria abundances and respiration rates could be related to the small differences in total bacterial counts observed between the different marine ecosystems investigated here. This could have masked the relationships between abundance and respiration rates.

While bacterioplankton abundance did not strongly vary, the structure of the community expressed by the percentage of HDNA cells was significantly different between sampling sites (Table 1). These values ranging between 61% and 85% are within the range of values observed in highly productive coastal marshes (Bouvier et al., 2007). This variability was primarily driven by the variability in LDNA cells abundances. Such shifts in bacterioplankton communities are mainly influenced by variability in nutrients and DOC (e.g. del Giorgio et al., 1997). In particular phosphate limitation of heterotrophic bacteria activity has been reported in coastal environments (e.g. Servais et al., 2003). The two sampling sites exhibiting the lowest percentage of active cells (rocky shore and seagrass bed) were characterized by low phosphate concentrations (Table 1). Assuming that LDNA bacteria are inactive or starved (Lebaron et al., 2002), then nutrient availability, and more particularly phosphate, could be a factor responsible for the variability in the LDNA sub-populations between sampling sites. However, results obtained across a wide range of aquatic ecosystems, have recently shown that the relative abundance of HDNA and LDNA cells within bacterioplankton communities was not only related to the overall system productivity and that other factors could influence the proportion of HDNA cells within bacterioplankton communities (Bouvier et al., 2007). A variety of bottom up and top down control factors not investigated here could have triggered the differences observed between our five sampling sites, e.g. variability in DOC concentration and composition (e.g. del Giorgio *et al.*, 1997), grazing by protozoans and viral infection (Fuhrman, 1999; Gasol *et al.*, 1999; Pedrós-Alió *et al.*, 2000; Bouvier *et al.*, 2007).

Relative abundance of HDNA cells and respiration rates

Our results highlighted a significant increase in respiration rates with the percentage of HDNA cells observed within the bacterioplankton communities (Figure 4). The cluster of bacterioplankton communities into HDNA and LDNA subpopulations has been observed over a wide range of aquatic ecosystems and appears to be a general character of bacterioplankton communities (e.g. Gasol & del Giorgio, 2000; Bouvier et al., 2007). Differences in metabolic activity between these two sub-populations have been suggested from field and laboratory work, with the HDNA cells representing the more active cells and LDNA corresponding with less active or dormant (Gasol et al., 1999; Lebaron et al., 2002). The positive correlation between respiration rates and the fraction of HDNA cells within the bacterioplankton community, tend to confirm this hypothesis. In addition, while the sheltered marina and the mudflat exhibited similar percentages of HDNA cells (respectively 85.5% and 83.8%; Table 1), respiration rates were significantly higher in the marina. This result suggests that the second HDNA population observed at the marina, exhibiting higher DNA content (Figure 2), may have higher metabolic rates that lead to the measurements of higher respiration rates. However, there is now evidence that low DNA cells may also be active (e.g. Zubkov et al., 2001; Longnecker et al., 2005). In addition there are conflicting results concerning the phylogenetic composition of these different fractions. Whilst some studies conclude that these fractions are phylogenetically different (e.g. Longnecker et al., 2005), others suggest that the composition varies between different fractions (e.g. Zubkov et al., 2001). Considering that HDNA and LDNA cells may correspond to two distinct and independent communities, as recent studies have shown a significant degree of phylogenetic dissimilarity between the groups (Zubkov et al., 2001; Mary et al., 2006), it is also realistic that different phylogenetic groups could exhibit divergent respiration rates. Recent evidence does not support the simple dichotomous view of HDNA 'active' and LDNA 'inactive' cells or the more extreme hypothesis of two distinct and independent communities, but rather suggests an intermediate scenario in which there are components intrinsic to both fractions, but also other components that switch between active and inactive (Bouvier et al., 2007).

The composition of bacterial communities was only defined here through their flow cytometric signature. Whilst our results suggest that the relative abundances of HDNA cells within the bacterioplankton, as well as the heterogeneity of HDNA sub-populations, should be taken into account when assessing the relation between bacterial community composition and its respiratory losses, future work is needed to conclude on the physiological and metabolic properties of the different bacterial sub-populations identified through flow cytometry.

Concluding remarks

Our study shows a strong connection between the fraction of HDNA cells within bacterioplankton communities and

respiration rates. More specifically, our results stress the need to relate bacterial respiration rates to the relative abundance of the different sub-populations found in bacterial assemblages. Assuming that only a fraction of bacterial cells contribute to bacterial respiration, the relative composition of bacterioplankton communities may have critical implications for calculations of bulk bacterial respiration and the related biogeochemical fluxes. However, the ecological role of the different cytometrically defined sub-populations within bacteria communities remained enigmatic and future work is needed to fully assess their role on biogeochemical processes.

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